

Rapid decline of the endemic giant land snail *Archachatina bicarinata* on the island of Príncipe, Gulf of Guinea

MARTIN DALLIMER and MARTIM MELO

Abstract Terrestrial and freshwater molluscs are amongst the most threatened of all taxa, yet data exist on the distribution and status of only a small proportion of the species. Here we present the results of the first systematic survey of a terrestrial mollusc on the island of Príncipe in the Gulf of Guinea. *Archachatina bicarinata* has never been previously surveyed, despite being categorized as Vulnerable on the IUCN Red List and suffering from unregulated harvesting for food. We found that *A. bicarinata* is restricted to primary rainforest and its abundance and probability of occurrence increased as surveyed sites became less accessible. Additional anecdotal evidence from the observations of previous scientific expeditions, local guides and snail harvesters suggested that the species has suffered a dramatic decline in population size and distribution in recent years. We therefore recommend that immediate action be taken to prevent its imminent extinction on Príncipe. The collection of *A. bicarinata* from protected areas should be banned, as should commercial harvesting and sales. Subsistence collecting should be limited to larger individuals. Our data have been used to inform the management plans for the protected areas on Príncipe and São Tomé, and this should help to ensure the future of *A. bicarinata* on both islands.

Keywords *Archachatina bicarinata*, bushmeat, decline, Gulf of Guinea, illegal harvesting, mollusc, over-exploitation, Príncipe

Introduction

Freshwater and terrestrial molluscs are highly threatened and have suffered the largest number of documented extinctions of any taxonomic group. However, only a small percentage of species (2% in 2004) have had their conservation status assessed (Lydeard et al., 2004). In many biodiversity studies molluscs are often ignored in favour of more charismatic fauna (Cameron, 1998; Lydeard et al., 2004), even though protected area designation for other taxa rarely adequately represents land snail diversity (Tattersfield, 1998;

Fontaine et al., 2007). In common with many other taxa, land snail diversity is affected by habitat change and land use (Raheem et al., 2008) and unregulated harvesting is a further threat to some terrestrial molluscs. Land snails are an important part of the local diet in parts of West Africa (Memel et al., 2008), where some people are full-time snail gatherers (Agbelusi & Ejidike, 1992). Organized trade in land snails has been recognized as a major reason for their decline in Nigeria, and availability for human consumption has been falling because of over-collection and habitat destruction (Osemeobo, 1992). Island mollusc faunas are particularly at risk and have suffered mass extinctions because of factors such as the introduction of invasive non-native species (Cowie & Cook, 2001; Coote & Loève, 2003).

The islands of São Tomé and Príncipe, in the Gulf of Guinea, are one of the 25 global biodiversity hot spots (Myers et al., 2000) and the archipelago is recognized as a Terrestrial Ecoregion by WWF (Gascoigne, 2004). Most research on the islands has focused on their birds (Jones & Tye, 2006; Dallimer et al., 2009) and vascular flora (Figueiredo, 1994). Nevertheless, there is also a high level of mollusc endemism. Príncipe alone has 19 single-island endemics (plus a further six species shared with just São Tomé) of 32 land snail species (Gascoigne, 1994a). Although the urgent need for distribution surveys and monitoring of the mollusc community was first highlighted in the mid 1990s (Gascoigne, 1994a), almost nothing is known about the conservation requirements of the various species. Of the 27 species of mollusc listed by the IUCN on São Tomé and Príncipe, only two have been given an IUCN status (IUCN, 2008). One of these is the giant land snail *Archachatina bicarinata* (endemic to São Tomé and Príncipe), which is categorized as Vulnerable solely because of its restricted range and the presumed threats from harvesting (Gascoigne, 1995). The species is an important part of the local diet on both islands, where it is preferred to *Archachatina marginata*, which was introduced from the African mainland, because of its larger size, superior taste and texture.

Here, we report the first systematic survey of the distribution of *A. bicarinata*. In addition, we collate anecdotal accounts of *A. bicarinata* abundance and distribution to document changes in population size and range.

Study area

Príncipe (area 139 km²) is the second largest oceanic island in the Gulf of Guinea. It lies 220 km from the African coast

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and 146 km north of São Tomé (Fig. 1). The north of the island is relatively flat and contains most of the island's human population and agriculture. To the south and centre the island is rugged and includes the highest mountain, Pico de Príncipe (948 m). Significant tracts of primary rainforest (33% of the island's total land area) remain in the south and west, which form the bulk of the Parque Natural Obô do Príncipe. However, the protected area has received little active conservation management and the forest is still subject to many of the pressures associated with unprotected areas, such as hunting and harvesting of forest products. Although it is believed that *A. bicarinata* is restricted to relatively undisturbed forest, no formal surveys of its distribution have previously been carried out. We therefore visited 13 sites, covering representative areas of the island, including primary and secondary forest and plantations (Fig. 2, Table 1).

Methods

At each site 4–10 150-m long, 2-m wide transects were sampled between 25 November and 20 December 2007 by scanning the forest floor and vegetation for both live and dead *A. bicarinata*. For many snail species the presence of empty shells is counted as evidence of their continued

existence (Raheem et al., 2008) and hence we also recorded these. Transects were at least 150 m apart and their position was recorded with a global positioning system where possible. Together with extensive field notes this allowed all transects to be located on a digital map of the island (Gascoigne et al., 2004) and relevant environmental information to be gathered. Habitat type (primary forest, secondary forest, plantations with shade canopy cover, palm plantation and agriculturally diverse) and climatic zone (1, very humid zone, typical total annual rainfall $\geq 3,000$ mm; 2, less humid zone, typical total annual rainfall c. 2,000 mm; 3, drier zone, typical total annual rainfall 1,000–1,500 mm) were classified following Diniz & de Matos (2002). Transect altitude and the relative productivity of the vegetation were assessed from satellite-derived remotely sensed data using a Digital Elevation Model and a measure of normalized difference vegetation index (NDVI), respectively. To quantify the accessibility of a site the distance to non-primary forest edge (which was negative for sites outside primary forest), distance to the nearest watercourse (as rivers often provide access routes into the forest) and distance to known settlements were also measured. Distance to the nearest road was not included in



FIG. 1 The location of Príncipe in the Gulf of Guinea, West Africa.

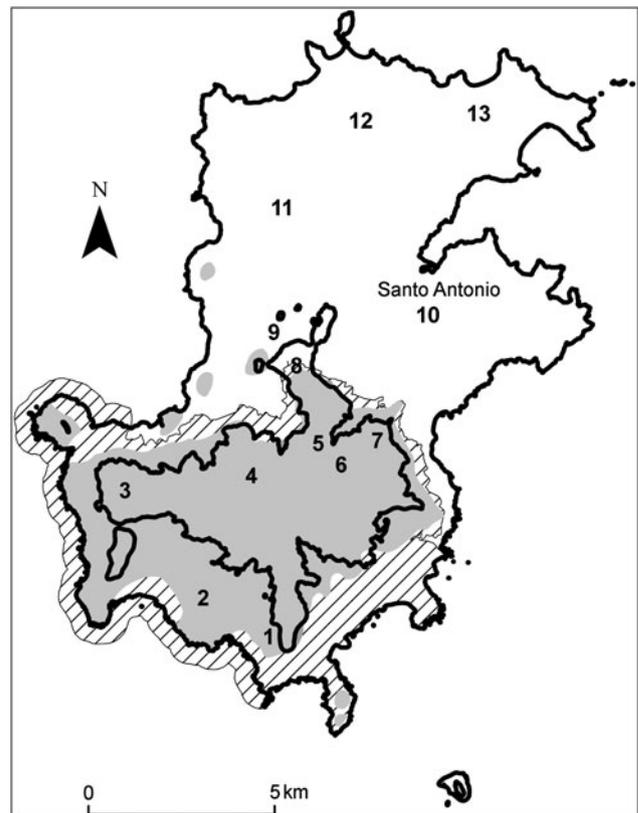


FIG. 2 The 13 survey sites (see Table 1 for further details) for *Archachatina bicarinata* on the island of Príncipe (Fig. 1). The grey-shaded area shows the extent of primary forest, the hatched area the extent of the Parque Natural Obô do Príncipe and the solid line represents the 250-m contour.

TABLE 1 Transect locations surveyed for *Archachatina bicarinata* on the island of Príncipe (Fig. 1). Locations of the 13 numbered transects are shown on Fig. 2.

| No. | Site name | Dominant habitat type | No. of transects surveyed | No. of transects with snails present | No. of snails per transect (range) |
|-----|-------------------|-----------------------|---------------------------|--------------------------------------|------------------------------------|
| 1 | Rio Porco | Primary forest | 10 | 2 | 0–1 |
| 2 | Camp Tomé | Primary forest | 10 | 4 | 0–2 |
| 3 | A Mesa | Primary forest | 5 | 5 | 2–9 |
| 4 | Pico do Príncipe | Primary forest | 9 | 6 | 0–1 |
| 5 | Boca do Inferno | Primary forest | 6 | 0 | 0 |
| 6 | O Que Pipi | Primary forest | 10 | 5 | 0–9 |
| 7 | Camp Joaquim | Secondary forest | 5 | 0 | 0 |
| 8 | Pico Papagaio | Secondary forest | 6 | 0 | 0 |
| 9 | Morro Estanduarte | Secondary forest | 4 | 0 | 0 |
| 10 | Bela Vista | Plantation | 5 | 0 | 0 |
| 11 | Ponta do Sol | Plantation | 5 | 0 | 0 |
| 12 | Ribeira Izé | Plantation | 5 | 0 | 0 |
| 13 | Belo Monte | Plantation | 5 | 0 | 0 |

any analyses as it is a strong covariate of distance to settlements.

We used multiple regression, with both the abundance of live snails (snail abundance) and the occurrence of live or dead individuals (snail presence) as the response variables, and measures of habitat type, productivity and location as explanatory variables. For snail abundance a generalized linear model with Poisson errors was used, and for snail presence a binomial error structure. The full model set therefore included habitat and climate types, NDVI, altitude, and distances to primary forest edge, watercourse and settlement. A quadratic term was included for altitude to model the potentially humped relationship with this variable. The information theoretic approach (Burnham & Anderson, 2002; Johnson & Ohmland, 2004) was used to model these data based on Akaike information criteria (AIC). All possible subsets of the variables were modelled. For the complete model set the difference in AIC for that model relative to the best fitting model with the minimum AIC (Δ AIC), the Akaike weight (w_i) and the proportion of the variation explained by the model (D^2) were calculated. The best fitting model was defined as that with the lowest AIC, and models that differed by less than two AIC units were considered to have substantial support in terms of explaining the data (Burnham & Anderson, 2002). The probability that each explanatory variable appeared in the best fitting model was also calculated (k). Parameter estimates and the relative contribution (partial D^2) of each explanatory variable to the overall predictive power of the model were calculated by model averaging across all models, weighting the estimates by w_i . All analyses were performed in R v. 2.7.2 (R Development Core Team, 2008).

While no formal surveys of terrestrial molluscs have previously been carried out on Príncipe, there are three indirect sources of data on recent population trends. Firstly,

at remote sites *A. bicarinata* are processed by snail harvesters before returning home. The presence of piles of snail shells (middens) may therefore give an indication of the impact of collectors and some impression of the size of *A. bicarinata* populations. Secondly, some researchers, while not specifically working on *A. bicarinata*, have recorded relevant observations. Thirdly, local guides and snail harvesters were asked to report their observations of *A. bicarinata* numbers and distribution. We were thus able to compare our systematic survey results with these three indirect sources of evidence on *A. bicarinata* abundance.

Results

Field surveys

A total of 85 transects were surveyed across 13 sites (Fig. 2). Six sites were completely within primary rainforest. No *A. bicarinata* were observed outside primary forest and therefore habitat type was not included in the analyses. Snails were found in 22 transects, with 45 (mean 0.53, range 0–9 per transect) live snails encountered and an additional 20 empty snail shells found.

For snail abundance, total model-averaged D^2 was high (0.58; Table 2). The best model contained three variables, indicating that more live individuals were found on transects (1) further away from the nearest settlement (partial $D^2 = 0.15$), (2) with higher productivity (NDVI; partial $D^2 = 0.23$) and (3) at intermediate altitudes as both linear and quadratic relationships with altitude were retained in the best model (Table 2).

Half the variation in snail presence ($D^2 = 0.50$) was predicted by the model-averaged parameters, with three variables appearing in the best model. Snails were more likely to be present further from the nearest settlements

TABLE 2 Model-averaged parameter estimates for habitat variables used in modelling snail abundance and presence on the island of Príncipe, Gulf of Guinea (Figs 1–2). Figures in bold indicate variables that appeared in the best model and those in italics indicate the additional variables that appeared in models with a $\Delta AIC < 2$.

| Variable | No. of live individuals | | | Evidence of snail presence | | |
|----------------------------|-------------------------|--------------------|---------------|----------------------------|--------------------|---------------|
| | <i>k</i> | Parameter estimate | Partial D^2 | <i>k</i> | Parameter estimate | Partial D^2 |
| Distance to nonforest | 0.35 | -35.91 | 0.09 | 0.39 | -75.82 | 0.08 |
| Distance to settlement | 1.00 | 115.96 | 0.15 | 0.99 | 219.79 | 0.16 |
| NVDI* | 0.96 | 0.00 | 0.23 | 0.61 | 0.00 | 0.06 |
| Distance to river | 0.37 | 64.05 | 0.00 | 0.68 | 494.38 | 0.01 |
| Climatic Zone 1 | 0.16 | | 0.00 | 0.18 | | 0.01 |
| Climatic Zone 2 | | 2.29 | | | 2.22 | |
| Climatic Zone 3 | | 2.37 | | | 2.51 | |
| Altitude | 1.00 | 0.03 | 0.02 | 0.62 | 0.01 | 0.11 |
| (Altitude) ² | 1.00 | 0.00 | 0.07 | 0.55 | 0.00 | 0.07 |
| Model-averaged total D^2 | | | 0.58 | | | 0.50 |

*Normalized difference vegetation index

(partial $D^2 = 0.16$) and at higher altitudes (partial $D^2 = 0.11$). Although predictive power was low (partial $D^2 = 0.01$), snail presence was more likely further away from watercourses.

Considering variables in the $\Delta AIC < 2$ model set, both snail abundance and likelihood of occurrence increased with greater distance from the forest edge, although predictive power was low (partial $D^2 = 0.09$ and 0.08 , respectively; Table 2). The only variable that did not appear in any model with a $\Delta AIC < 2$ was the climatic zone within which transects were situated.

Anecdotal accounts of past occurrence and abundance

Anecdotal evidence from scientific expeditions, local snail harvesters and guides was limited to the few locations that have harboured large populations of snails in the past. Nevertheless, all suggest that there has been a marked decline in *A. bicarinata* abundance and range in the past 5–10 years. For A Mesa (Site 3), the only place where snails were still abundant (Table 1), the presence of several middens containing many hundreds of snail shells suggested that populations are suffering heavy harvesting pressure. One midden contained an estimated 515 recognizable snail shells along with many others in a more advanced state of disintegration. Within the area covered by the transects there were a further two of similar size and two additional, smaller, middens containing 100–150 shells. Therefore, within this small area, c. 3,000 *A. bicarinata* individuals had been collected in recent years. Middens could potentially represent the build-up of shells over a long period of time. However, in this instance, the condition of many of the shells was similar, indicating that a only a few harvesting trips may have been responsible for the majority of the shells.

Traditionally, both Pico do Príncipe (Site 4) and Boca do Inferno (Site 5) have been visited by local snail harvesters as

these sites have supported large snail populations. At Boca do Inferno no live *A. bicarinata* were observed, and snail harvesters commented that they instead have to collect from deeper in the forest around Pico do Príncipe, where population sizes remain higher. One guide stated that in 2002 at Pico do Príncipe *A. bicarinata* was so common that they were ‘like stones’. This site has been visited twice by scientific expeditions in the past 10 years, both of which noted the abundance of *A. bicarinata*. In 2001, an expedition collected 20 *A. bicarinata* individuals on a single night for dinner (Californian Academy of Sciences, 2001). Baillie (1999) was ‘struck by the abundance of giant snails. At one point we were finding snails at one metre intervals.’ In contrast, we observed only a single live snail (Table 1).

Discussion

A. bicarinata is both more likely to occur and be more abundant at sites that are less accessible and at higher altitude. A similar pattern has been observed for the globally threatened birds of Príncipe (Dallimer & King, 2008). *A. bicarinata* does not occur in human-modified landscapes. The snail is therefore harvested both opportunistically when people are in the forest and during specific collecting trips, both for personal consumption and commercial sale. The small size of Príncipe means that all the forest is within only a few km of the nearest entry points. Even snail populations at the more remote sites are therefore vulnerable to collection pressure from harvesters, local guides and research scientists. Nevertheless, the less accessible forest has acted as a refuge, allowing populations of *A. bicarinata* to persist.

As no previous systematic surveys of *A. bicarinata* have taken place, our data provide a baseline for future monitoring. Although we are only able to present anecdotal evidence of population decline the records indicate a population

collapse. More accessible parts of the forest (e.g. Boca do Inferno) have lost almost their entire snail population and areas deep inside the forest, where snails were once abundant, now have markedly smaller populations (e.g. Pico do Príncipe and A Mesa). For other species seasonal changes in detectability can make comparing survey results from different times of the year problematic. Recent scientific expeditions to Pico do Príncipe took place in September (Baillie, 1999) and May (California Academy of Sciences, 2001), whereas our surveys were in December. Although it remains possible that *A. bicarinata* undergoes seasonal changes either in abundance or in detectability associated with its annual life cycle, it is unlikely that these could be responsible for the dramatic decline reported here.

Two species of snail are routinely collected for food on São Tomé and Príncipe. The introduced *A. marginata* is common throughout the agricultural areas of the island and never encountered in the forest, whereas *A. bicarinata* is restricted to primary forest. Formerly, however, when the extent of forest was greater, *A. bicarinata* may have occurred across a larger portion of the islands (Gascoigne, 1994b). *A. marginata* is a recent arrival on Príncipe (*A. Gascoigne, pers. comm.*), as it was not present in the early 1990s (Gascoigne, 1994a,b), and has spread rapidly throughout the human-modified environments within the past 20 years. It is unknown what impact this introduction may have had on the distribution of *A. bicarinata*, and it is possible that *A. marginata* may have brought over an infectious disease or parasite to which *A. bicarinata* has no resistance (Gascoigne, 1994b).

In common with many terrestrial molluscs in island systems *A. bicarinata* is severely threatened on Príncipe. Although no formal surveys have been carried out on São Tomé, fieldworkers have reported a similar situation there and hence it is thus possible that *A. bicarinata* could rapidly become extinct. Even if the population collapse were limited to Príncipe, immediate action is required if the species is not to disappear from the island.

Príncipe currently retains a relatively large area of primary rainforest. The recent legal designation of this as a protected area could help secure *A. bicarinata* and, although management resources are limited, the forest covers a substantial region within which the conservation of the remaining populations can be prioritized. However, we recommend, firstly, that harvesting of snails from protected areas on both Príncipe and São Tomé is immediately banned (although enforcing the ban is likely to be difficult and would need to be carried out in conjunction with an awareness campaign). Secondly, the commercial collection and sale of *A. bicarinata* should be made illegal. Thirdly, harvesting for subsistence purposes should be limited to large individuals only; the precise measurements of such snails would need to be clarified by further research. If populations recover, some relaxation of these restrictions

may be possible but this should only be done within the context of a good understanding of population dynamics and sustainable harvesting rates, possibly by implementing an adaptive management approach (McCarthy & Possingham, 2006). Fourthly, we appeal to scientists, ornithologists and tourists visiting these islands not to eat this endemic species and to express their concerns to guides and tour operators if they observe that snails are being collected for human consumption.

Following our survey the Department of Protected Areas for São Tomé and Príncipe has taken some action to promote the conservation of *A. bicarinata*. As part of the management plan for the protected areas currently being drafted by the government and ECOFAC (Programme de Conservation et Utilisation Rationnelle des Ecosystemes Forestiers en Afrique Centrale; a European Union-funded conservation programme for the forests of central Africa), *A. bicarinata* was chosen as one of a suite of indicator species that will be regularly surveyed to assess the effectiveness of the protected areas for biodiversity conservation. In addition, the NGO Associação Monte Pico will start work on Príncipe with snail harvesters, with the aim of re-training them as guides and support staff for the protected area.

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References

- AGBELUSI, E.A. & EJIDIKE, B.N. (1992) Utilization of the African giant land snail *Archachatina marginata* in the humid area of Nigeria. *Tropical Agriculture*, 69, 88–92.
- BAILLIE, J. (1999) *One Month in the Forests of Príncipe*. [Http://www.ggcg.st/jon_principe.htm](http://www.ggcg.st/jon_principe.htm) [accessed 5 January 2009].
- BURNHAM, K.P. & ANDERSON, D.R. (2002) *Model Selection and Multimodal Inference*. Springer, New York, USA.
- CALIFORNIA ACADEMY OF SCIENCES (2001) *Californian Academy of Sciences Gulf of Guinea Biodiversity Project*. [Http://www.calacademy.org/science_now/sao_tome/dispatch10.htm](http://www.calacademy.org/science_now/sao_tome/dispatch10.htm) [accessed 5 January 2009].

- CAMERON, R.A.D. (1998) Dilemmas of rarity: biogeographical insights and conservation priorities for land Mollusca. *Journal of Conchology*, Special Issue, 51–59.
- COOTE, T. & LOËVE, É. (2003) From 61 species to five: endemic tree snails of the Society Islands fall prey to an ill-judged biological control programme. *Oryx*, 37, 91–96.
- COWIE, R.H. & COOK, R.P. (2001) Extinction or survival: partulid tree snails in American Samoa. *Biodiversity and Conservation*, 10, 143–159.
- DALLIMER, M. & KING, T. (2008) Habitat preferences of the forest birds on the island of Príncipe, Gulf of Guinea. *African Journal of Ecology*, 46, 258–266.
- DALLIMER, M., KING, T. & ATKINSON, R.J. (2009) Pervasive threats within a protected area: conserving the endemic birds of São Tomé, West Africa. *Animal Conservation*, 12, 209–219.
- DINIZ, A.C. & DE MATOS, G.C. (2002) Carta de Zonagem Agro-Ecológica e da Vegetação de S. Tomé e Príncipe: I Ilha de S. Tomé. *Garcia de Orta, Ser Botanica*, 15, 1–22.
- FIGUEIREDO, E. (1994) Diversity and endemism of angiosperms in the Gulf of Guinea islands. *Biodiversity and Conservation*, 3, 785–793.
- FONTAINE, B., GARGOMINY, O. & NEUBERT, E. (2007) Priority sites for conservation of land snails in Gabon: testing the umbrella species concept. *Diversity and Distributions*, 13, 725–734.
- GASCOIGNE, A. (1994a) The biogeography of land snails in the islands of the Gulf of Guinea. *Biodiversity and Conservation*, 3, 794–807.
- GASCOIGNE, A. (1994b) The dispersal of terrestrial gastropod species in the Gulf of Guinea. *Journal of Conchology*, 35, 1–7.
- GASCOIGNE, A. (1995) *A Lista Vermelha de Animais Ameaçados de São Tomé e Príncipe*. ECOFAC, São Tomé, São Tomé and Príncipe.
- GASCOIGNE, A. (2004) São Tomé, Príncipe, and Annobon moist lowland forests. In *Terrestrial Ecoregions of Africa and Madagascar: A Conservation Assessment* (eds N. Burgess, J. D'Amico Hales, E. Underwood & E. Dinerstein), pp. 236–238. Island Press, Washington, DC, USA.
- GASCOIGNE, A., WOJCIECHOWSKI, T. & KOO, M. (2004) *Gulf of Guinea Conservation Group: Maps*. <http://www.ggcg.st/maps/mapintro.html> [accessed 16 November 2008].
- IUCN (2008) *2008 IUCN Red List of Threatened Species*. <http://www.iucnredlist.org> [accessed 5 January 2009].
- JOHNSON, J.B. & OHMLAND, K.S. (2004) Model selection in ecology and evolution. *Trends in Ecology & Evolution*, 19, 101–108.
- JONES, P.J. & TYE, A. (2006) *The Birds of São Tomé and Príncipe with Annobón, BOU Checklist Series No. 22*. British Ornithologists Union, Oxford, UK.
- LYDEARD, C., COWIE, R.H., PONDER, W.F., BOGAN, A.E., BOUCHET, P., CLARK, S.A. et al. (2004) The global decline of non-marine mollusks. *BioScience*, 54, 321–330.
- MCCARTHY, M.A. & POSSINGHAM, H.P. (2006) Active adaptive management for conservation. *Conservation Biology*, 21, 956–963.
- MEMEL, J.D., OTCHOUMOU, A., KOUASSI, D.K. & DOSSO, H. (2008) Inventaire, potentiel et répartition des escargot terrestres d'une forêt tropicale humide de Côte d'Ivoire: le Parc National du Banco (PNB). *Novapex*, 9, 119–127.
- MYERS, N., MITTERMEIER, R.A., MITTERMEIER, C.G., DA FONSECA, G.A.B. & KENT, J. (2000) Biodiversity hotspots for conservation priorities. *Nature*, 403, 853–858.
- OSEMEOBO, G.J. (1992) Effects of land-use and collection on the decline of African giant snails in Nigeria. *Environmental Conservation*, 19, 153–159.
- R DEVELOPMENT CORE TEAM (2008) *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna, Austria. <http://www.R-project.org> [accessed January 2009].
- RAHEEM, D.C., NAGGS, F., PREECE, R.C., MAPATUNA, Y., KARIYAWASAM, L. & EGGLETON, P. (2008) Structure and conservation of Sri Lankan land-snail assemblages in fragmented lowland rainforest and village home gardens. *Journal of Applied Ecology*, 45, 1019–1028.
- TATTERSFIELD, P. (1998) Patterns of diversity and endemism in East African land snails, and the implications for conservation. *Journal of Conchology*, Special Issue, 77–86.

Biographical sketches

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