# Long-term community-based monitoring of tamaraw *Bubalus mindorensis* on Mindoro Island, Philippines

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**Abstract** The Critically Endangered tamaraw Bubalus mindorensis is endemic to Mindoro Island, Philippines, and little is known of its ecology. During 2006-2011 we used community-based monitoring to examine the population status and ecology of tamaraw in the species' core habitat of Mount Iglit-Baco National Park. Each year, for 5 consecutive days at the end of the dry season, trained local volunteers and rangers or project staff were allocated to 18 vantage points in the study area (c. 160 km²). Tamaraw were categorized as adult (> 5 years), juvenile (2-5 years) or calf (< 2 years), and sexed when possible. During the study period the population was 239-314 (mean 271), with no significant fluctuations in age structure (percentage of adults, juveniles and calves: 57.8, 21.0 and 21.3%) or estimated adult female reproductive rates (29.1%). In adults, but not in juveniles, the sex ratio was biased towards females (1:1.86, P < 0.01). Bulls were often solitary (32.2% of sightings), whereas the majority of cows (94.7%) formed small groups of 2-12 individuals of different ages, with or without bulls (53.4 and 46.6%, respectively). These results demonstrate that the population remained relatively stable, maintaining a constant age structure and reproductive rate, and that long-term community-based monitoring was

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Received 11 December 2012. Revision requested 7 June 2013. Accepted 29 April 2014. First published online 1 August 2014. effective for quantitative characterization of the tamaraw's social behaviour, which is critical for conservation and management of the species.

**Keywords** *Bubalus mindorensis*, community-based monitoring, Critically Endangered species, Mindoro Island, population trend, tamaraw

#### Introduction

The Philippine archipelago comprises a land area of L c. 300,000 km<sup>2</sup> and hosts 206 native species of land mammals. It is a biodiversity hotspot but 117 (57%) of the endemic native land mammals are threatened by habitat loss, illegal hunting and human encroachment (Myers et al., 2000). The tamaraw *Bubalus mindorensis*, which is endemic to the island of Mindoro, is categorized as Critically Endangered on the IUCN Red List (Hedges et al., 2008). It is believed that tamaraw were once distributed throughout the island, coexisting with the indigenous Mangyan people, who traditionally practised slash-and-burn agriculture (Everett, 1878; Steere, 1888). However, the tamaraw population decreased during the early 20th century as a result of increased human activity, particularly hunting (Manuel, 1957; Harrisson, 1969), and is currently estimated to be c. 300. Its distribution is restricted to three protected areas (Mount Iglit-Baco National Park, Mount Aruyan-Malati Tamaraw Reservation and Mount Calavite Wildlife Sanctuary), with > 90% of individuals belonging to the sub-population of Mount Iglit-Baco National Park (Hedges et al., 2008).

Information on population status is essential for wild-life conservation and management (Primack, 1993), and an understanding of ecological traits such as social behaviours, mating systems and habitat selection is necessary to design a successful wildlife conservation programme (Curio, 1996; Caro, 1999; Anthony & Blumstein, 2000). However, for Critically Endangered species, particularly in tropical countries, such information is not widely available in the scientific literature (Collen et al., 2008). Some studies have examined the physiological or cytogenetic traits of tamaraw in captivity but little is known about the ecology of tamaraw in the wild (Maala, 2001).

The scarcity of data from tropical countries is attributed to inappropriate resource allocation (Sheil, 2001) or insufficient funding and lack of adequate infrastructure and expertise for data collection (Collen et al., 2008). To address this situation it has been suggested that monitoring of threatened species may benefit from a participatory approach (Danielsen et al., 2009) and that local community-based monitoring can be as reliable as professional monitoring if it is performed with attention to objectives, methodological designs and intensive analyses (Burton, 2012).

In 1999 the Department of Environment and Natural Resources of the Philippines shifted the priority of its Tamaraw Conservation Programme from captive breeding to the protection of wild populations in Mount Iglit-Baco National Park in accordance with the recommendations of the Population and Habitat Viability Assessment for Tamaraw (IUCN/SSC Conservation Breeding Specialist Group, 1996). The Tamaraw Conservation Programme is linked with the Community Assistance Programme, which is a livelihood assistance programme for local communities within or near tamaraw habitats, and the nationwide Information, Education and Communication Campaign, which aims to educate people, organize local partnerships and develop local biodiversity conservation. A large-scale community-based monitoring programme, the Tamaraw Population Count, was initiated to determine the number of tamaraw in the core habitat in Mount Iglit-Baco National Park (Philippine Clearing House Mechanism, 2012) and this has become a regular annual activity since 2000, providing realistic estimates of the tamaraw population in the Park (Department of Environment and Natural Resources Region 4B, 2012). This suggests that community-based tamaraw monitoring can generate relevant ecological data and population estimates, with only minor modifications in data collection and analysis procedures.

Our objective was to use community-based monitoring to improve our understanding of tamaraw ecology in the wild. Data on the ecology of tamaraw were collected in the Mount Iglit range over 6 years and were used to estimate population size, age structure, reproductive rates of adult females and social behaviours. Data on the prevalence of endoparasites were also collected, to evaluate the general health of wild tamaraw.

#### Study area

Mount Iglit–Baco National Park (754 km²) is located in central Mindoro Island (Fig. 1). It has a tropical monsoon climate, with dry (December–May) and wet (June–November) seasons. The tamaraw conservation area in the Park is c. 160 km² and comprises hilly terrain at altitudes of

300–1,000 m. High-altitude areas are dominated by perennial grasses such as cogon *Imperata cylindrica*, talahib *Saccharum spontaneum* and themeda *Themeda triandra*, whereas low-altitude areas are characterized by grassland, with patches of secondary forest and bamboo; the vegetation along rivers is sufficiently dense to provide cover for tamaraw (Maala, 2001; Cebrian et al., 2014). The Philippine deer *Rusa marianna* and Oliver's warty pig *Sus oliveri* have also been observed in the area but no tamaraw predators have been reported.

#### Methods

Participants in the Tamaraw Population Count included schoolteachers, public workers, NGOs and local government employees. Prior to the field surveys these volunteer observers (20–30 people) were trained in survey methods, including protocols for counting, identification of age and sex, and use of binoculars, maps and compasses. They were allocated to one of 18 vantage points in the tamaraw conservation area, where at least one ranger or project staff member was concurrently assigned, to minimize possible counting errors. There were c. 50 observers in total.

Tamaraw were counted simultaneously at the 18 vantage points each April during 2006-2011 (IUCN/SSC Conservation Breeding Specialist Group, 1996; Fig. 1c). Observers were equipped with compasses and binoculars and/or spotting scopes. Controlled burning of grass was carried out 1-2 months before the count, to facilitate visibility and to entice tamaraw into the open with an abundance of emerging sprouts and young grasses; burning reduced the total grass biomass temporarily but increased the ratio of the edible portion of grasses and its crude protein content significantly at the time of the tamaraw count (R. Lapitan et al., unpubl. data). To minimize disturbance to the tamaraw burning was limited to the main grasslands and was carefully controlled to prevent the fire from spreading to forest edges or river basins, which tamaraw use as resting places and travel corridors.

Observations were made simultaneously at all vantage points for 5 consecutive days (22–26 April) at 05.30–07.00 and 17.00–18.30, when tamaraw generally graze (Momongan & Walde, 1993). For each sighting the time and the direction and distance of the tamaraw from the observation site were recorded. When groups of tamaraw were sighted, group size and composition were recorded.

Tamaraw were categorized as adults (> 5 years, sexually mature), juveniles (2-5 years) or calves (< 2 years), according to previous studies (Kuehn, 1986; Ong et al., 1996), based on body conformation, fur colour and horn size: calves, reddish brown or light brown fur, no horns or horns not reaching the ear tips; juveniles, obvious black

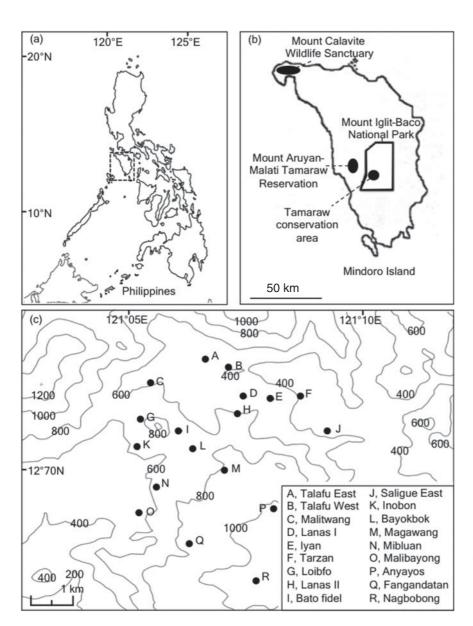


Fig. 1 (a) Location of Mindoro Island in the Philippines, (b) habitat ranges of tamaraw *Bubalus mindorensis* on the island, and (c) topography of the tamaraw conservation area (c. 160 km²) in Mount Iglit–Baco National Park (754 km²), and locations of the 18 vantage points used in the survey. Tamaraw have been confirmed in three protected areas on Mindoro, with > 90% of individuals inhabiting the tamaraw conservation

dorsal stripe, spike-like horns extending beyond ear tips; adults, brown to greyish-black fur with pale dorsal stripes, v-shaped horns extending over the head.

The sex of adult or juvenile tamaraw was determined based on body conformation, horn shape and genitalia when visible. Adult males have wider and thicker horns than females and are more muscular. When it was difficult or impossible to determine the sex of individuals they were categorized as sex-unidentified.

The prevalence of endoparasites was investigated as an indicator of tamaraw health status. Fresh faecal samples from tamaraw were collected in the vicinity of observation sites and stored in 10% formalin. Subsequently, endoparasite eggs were detected using a sedimentation method (Hendrix, 1998).

Data from each observation site were consolidated during a workshop after the 5-day field survey each year

and any possible double or multiple counts from two or more observation sites were removed. The sex ratios among adults and juveniles were calculated, excluding sexunidentified animals. The reproductive rate of adult females was estimated from the ratio between the number of calves <1 year old and the number of adult females. In this estimation half of the animals categorized as calves were assumed to be <1 year old and half of the sex-unidentified adults were assumed to be females and were included in the total adult female population.

Binomial tests were used to determine whether observed sex ratios differed from the expected sex ratio of 1:1. We used  $\chi^2$  tests to determine whether the distribution of categorical variables such as gender and age were independent of the year. Statistical differences between means were examined using a Student's *t*-test. Differences were considered significant when P < 0.05.

TABLE 1 Annual total count, age structure, and estimated sex ratio of the tamaraw Bubalus mindorensis population in Mount Iglit-Baco National Park, Mindoro Island, Philippines (Fig. 1), during 2006-2011.

		Age structure <sup>1</sup>			Sex ratio of adults <sup>2</sup>	adults <sup>2</sup>		Sex ratio of juveniles <sup>2</sup>	juveniles <sup>2</sup>	
		Adults,	Juveniles,	Calves,	Males	Females		Males	Females	Sex
Year	Total count	> 5 years (%)	2–5 years (%)	0-2 years (%)	identified	identified	Sex ratio	identified	identified	ratio
2006	263	162 (61.6)	52 (19.8)	49 (18.6)	53	102	$1:1.92^{**}$	8	12	1:1.50
2007	239	132 (55.2)	48 (20.1)	59 (24.7)	53	29	$1:1.49^{*}$	13	20	1:1.53
2008	263	155 (58.9)	66 (25.1)	42 (16.0)	49	86	$1:2.00^{**}$	16	6	1:0.56
2009	274	151 (55.1)	51 (18.6)	72 (26.3)	49	66	$1:2.02^{**}$	6	12	1:1.33
2010	314	172 (54.8)	74 (23.6)	68 (21.6)	09	108	$1:1.80^{**}$	23	19	1:0.82
2011	274	166 (60.6)	51 (18.6)	57 (20.8)	50	66	$1:1.98^{**}$	13	12	1:0.92
Mean or total <sup>3</sup>	$271.2 \pm SE 10.0$	$156.3 \pm SE 5.7 (57.7)$	$57.0 \pm SE 4.2 (21.0)$	$57.8 \pm SE 4.6 (21.3)$	314	585	1:1.86**	82	84	1:1.02

Ratios of adults, juveniles and calves did not differ between years ( $\chi^2 = 15.945$ , df = 10). Sex-unidentified individuals were excluded from calculations of sex ratios. Mean values over 6 years for total count and age structure; total count over 6 years for the sex ratio of adults and juveniles.

Table 2 Estimation of reproductive rates among a dult females in the tamaraw population in Mount Iglit–Baco National Park (Fig. 1) during 2006–2011.

Year	No. of adult females <sup>1</sup>	No. of calves < 1 year old <sup>2</sup>	Estimated adult female reproductive rate <sup>3</sup> (%)
2006	105.5	24.5	23.2
2007	79	29.5	37.3
2008	102	21	20.6
2009	100.5	36	35.8
2010	110	34	30.9
2011	107.5	28.5	26.5
Mean	$100.8\pm SE4.6$	$28.9 \pm SE\ 2.3$	$29.1 \pm SE 2.8$

'Half of sex-unidentified adult animals were considered female and included in adult female population count.

#### **Results**

The number of tamaraw counted during each year of the study was 239–314 (mean 271), with a small fluctuation (coefficient of variation 0.09) over the 6 years (Table 1). On average, the percentages of adults, juveniles and calves were 57.7, 21.0 and 21.3%, respectively. The age structure did not differ from year to year but the number of juveniles per 1 year of age was significantly lower than that of calves over the 6-year period (19.0  $\pm$  SE 1.4 vs 28.9  $\pm$  SE 2.3; Student's t-test, P < 0.05).

Sex identification was possible for 96% of adults and 49% of juveniles (Table 1). The sex ratio of adults was consistently biased towards females (overall adult male: female ratio = 1:1.86; binomial test, z=9.038, P < 0.01) but no bias was observed among juveniles. Reproductive rates among adult females were 20.6–37.3% (mean 29.1%), and did not differ significantly between years (Table 2).

Of the 1,627 tamaraw counted, 10% were alone and 90% were in groups (Table 3). Although there was a consistent tendency towards group formation, the proportion of solitary animals varied annually ( $\chi^2 = 20.453$ , df = 5, P < 0.01). Of the 163 solitary animals 101 (61.9%) were adult males, 31 (19.0%) were adult females, nine (5.5%) were sex-unidentified adults, and others included 19 juveniles and three calves. Groups comprised 2–12 individuals (overall mean of 3.57); the majority of the groups (88.9%) comprised 2–5 individuals.

Most groups included one or two adult females (358/378,94.7%) and approximately half of the groups (200/378, 52.9%) were accompanied by an adult bull (Table 4). Tamaraw in groups stayed within c. 10 m of each other and almost all groups remained in the same location during 5 consecutive days of observation. They were often observed

 $<sup>^2</sup>$ Half of animals identified as calves in Table 1 were assumed to be  $\leq$  1 year old

<sup>&</sup>lt;sup>3</sup>Estimated adult female reproductive rates did not differ by year ( $\chi^2 = 5.923$ , df = 5).

Table 3 The number of tamaraw observed alone or in groups in Mount Iglit-Baco National Park (Fig. 1) during 2006-2011.

	No. of solitary	No. of tamaraw in	Total no. of	Mean group	No. o	f groups	of 2-	12 tam	araw			
Year	tamaraw* (%)	groups* (%)	groups	size	2	3	4	5	6	7	8	12
2006	29 (11.0)	234 (89.0)	71	3.30	23	21	17	6	1	3	0	0
2007	12 (5.0)	227 (95.0)	59	3.85	12	13	16	12	2	4	0	0
2008	19 (7.0)	244 (93.0)	62	3.94	14	15	12	10	5	5	1	0
2009	26 (9.5)	248 (90.5)	72	3.44	24	22	10	7	5	3	1	0
2010	33 (10.5)	281 (89.5)	73	3.85	21	19	11	7	8	5	1	1
2011	44 (16.1)	230 (83.9)	76	3.03	31	23	13	8	0	1	0	0
Overall	163 (10.0)	1,464 (90.0)	413	3.57	125	113	79	50	21	21	3	1

<sup>\*</sup>Ratios of solitary animals and non-solitary animals differed significantly by year ( $\chi^2 = 20.453$ , df = 5, P < 0.01).

Table 4 Composition of tamaraw groups in terms of numbers of adult males and females.\*

No. of adult	No. of adult males in a group							
females in a group	0	1	2	Overall				
1	124 (32.8%)	109 (28.8%)	1 (0.3%)	234 (61.9%)				
2	45 (11.9%)	78 (20.6%)	1 (0.3%)	124 (32.8%)				
3-4	7 (1.9%)	13 (3.4%)	0 (0.0%)	20 (5.3%)				
Overall	176 (46.6%)	200 (52.9%)	2 (0.5%)	378 (100.0%)				

<sup>\*</sup>Groups lacking an identified adult female are excluded (35 of 413 in Table 3).



PLATE 1 A typical tamaraw *Bubalus mindorensis* herd comprising one adult male (upper middle), two adult females (1st right and 2nd from the left), one juvenile (lower middle) and one yearling (1st left), near the Loibfo vantage point in 2007 (photographed by Hisashi Matsubayashi).

emerging from the forest or creek areas onto the grassland at sunset and were hidden in forested areas or in tall grasses soon after sunrise. Plate 1 shows a typical tamaraw group comprising five individuals.

Two cases of intense fighting between bulls were witnessed during the study. In the first case the defeated bull suffered a serious injury and in the second case the defeated bull fell over a cliff. In addition, the carcass of a young bull was detected near the Magawang vantage point in 2011, in which a clear puncture wound apparently sustained in a fight was observed on the neck.

Endoparasites were detected in six of 27 (22%) fresh faecal samples but there were no superinfections. The genera were *Fasciola* (n = 2), *Eimeria* (n = 2), *Strongylus* (n = 1) and *Ascaris* (n = 1).

#### Discussion

Population trends and general ecology

Total counts and annual fluctuations were similar to those recorded during 2002–2005 (Department of Environment and Natural Resources Region 4B, 2012). Our results indicate a stable tamaraw population in Mount Iglit–Baco National Park and sustained precision in community-based tamaraw counting by the Tamaraw Conservation Programme. However, the total counting method generally provides a minimum live estimate because it is unlikely that all individuals will be observed (WWF, 2004), particularly given the tamaraw's nocturnal grazing patterns (Kuehn, 1986). Moreover, animals entering and leaving the study area cannot be ruled out because the area is not a closed system, and therefore a larger study area could result in a higher population estimate, although there is currently no evidence that tamaraw live outside the study area.

Age structure and estimated reproductive rates of adult females did not vary significantly between years, suggesting that the incidence of births and deaths was balanced. Our estimates did not account for calf mortality or that calves may have been concealed in the undergrowth, and therefore reproductive rates could be higher than estimated. The estimated calf: cow ratio (0.29) was lower than that reported for closely related species, such as African buffalo *Syncerus caffer* (0.70; Sinclair, 1977), American bison *Bison bison* (0.62; Van Vuren & Bray, 1986) and wild gaur *Bos gaurus* (0.34; Ahrestani et al., 2011), but similar to that reported for wild water buffalo *Bubalus arnee* in Nepal (0.29; Heinen, 1993), in which a decline in the calf: cow ratio has been reported (Khatri et al., 2012). As female reproductive rates are a key parameter of population viability, further investigation is required to determine whether the reproductive rates recorded by the monitoring are within the intrinsic performance range of this species or are suppressed by threats in the environment, such as nutrient deficiency as a result of habitat loss.

The lower abundance of juveniles per 1 year of age than of calves suggests that calf mortality is probably high, as reported for other ungulate species (Spinage, 1972). Tamaraw have no obvious predators, and therefore we could assume that shortage of food or resultant malnutrition were the main cause of death. However, most animals had good body conformation and levels of endoparasitic infection were lower than those observed in captive tamaraw (Anunciado et al., 1994). Malnutrition and endoparasitic load are therefore less likely to be major causes of tamaraw mortality, at least in the dry season. However, it is necessary to examine the health status of tamaraw further because endoparasitic infections are generally more prevalent during the wet season (Fritsche et al., 1993; Kuchai et al., 2011) and little is known about habitat use by tamaraw.

The female-biased adult sex ratio observed here is in agreement with observations in other sexually dimorphic ungulate species (African buffalo: Sinclair, 1977; kudu *Tragelaphus strepsiceros*: Owen-Smith, 1993), among which adult males are believed to suffer higher mortality than females because of higher growth rates, intense intra-male competition and solitary habits (Emlen & Oring, 1977; Berger & Gompper, 1999). Observations of intense fights between bulls and the carcass of a young bull with a neck wound suggest that such reasoning may also apply to tamaraw.

Tamaraw are often described as solitary (Talbot & Talbot, 1966; Kuehn, 1976; Suchomel, 2005). Following 221 days of observation across the wet and dry seasons Kuehn (1986) reported that 82% (179/218) of bulls were solitary individuals and 66% (71/107) of cows were alone or with calves. By contrast, here we observed that 32% (101/314) of bulls were solitary and 95% (554/585) of cows were in small family groups. Social segregation of adult males and females is often observed during the non-breeding season in many ungulate species of temperate regions (Ruckstuhl & Neuhaus, 2002). However, some tropical ungulate species, such as the African buffalo, adopt periodical segregation throughout the extended breeding season (Turner et al., 2005). Because the tamaraw breeding season and social

system are poorly known we cannot infer sexual segregation from our study.

# Efficacy of community-based tamaraw monitoring

Local community-based monitoring is expected to elucidate the status of species and habitats in tropical countries where professional monitoring may be difficult or ineffective, but the reliability of the data collected has been questioned (Danielsen et al., 2005). The accuracy and precision of such monitoring programmes vary according to relative contributions from local stakeholders and professional researchers: larger contributions from local people have resulted in lower reliability but empowered local communities to manage their resources more effectively (Danielsen et al., 2009). Local community-based monitoring has underestimated the size of mammalian communities but has provided reliable population estimates, similar to those of professional monitoring, for large ungulates (Burton, 2012). Knowledge of the strengths and weaknesses of participatory approaches can be used to maximize the potential of community-based monitoring. The small variations in total counts and age-sex structures recorded here are probably attributable to the well-prepared counting procedures, such as the allocation of skilled staff in addition to local volunteers at each observation site, and execution of scheduled burning to facilitate sightings, although the latter may have influenced habitat use by tamaraw to some extent.

Human activities, particularly hunting and exploitation, have the greatest effect on the population dynamics of wild animals (Redford, 1992; Mace & Reynolds, 2001). The Philippine government prohibited the hunting of tamaraw as early as 1936 but poaching did not stop completely. A field survey in the Mount Aruyan-Malati Tamaraw Reservation revealed that few tamaraw remain there, and they are threatened by poaching and land-use conflicts with the Mangyan people (Matsubayashi et al., 2010). A similar field survey at Mount Calavite Wildlife Sanctuary in 2010 failed to find evidence of the presence of tamaraw (S. Ishihara et al., unpubl. data). Despite the critical situation in these two areas, which had been regarded as potential tamaraw habitats, the tamaraw population appears to be stable in the Mount Iglit range of Mount Iglit-Baco National Park. In addition to population counting, community-based conservation activities such as the Community Assistance Programme and the Information, Education and Communication Campaign may deter poaching and safeguard against habitat destruction and human-tamaraw conflict. Our results indicate a need for continued protection of tamaraw habitat in Mount Iglit-Baco National Park, continued community-based monitoring, more research into tamaraw ecology and increased

public awareness and participation in tamaraw conservation.

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# **Biographical sketches**

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